

Title:

Forecasting the response of Great Lakes sea lamprey (*Petromyzon marinus*) to barrier removals

Authors:

Alexander James Jensen¹

jensen.alex1502@gmail.com

Michael L. Jones¹

jonesm30@msu.edu

Author Affiliation/Address:

¹Michigan State University, Department of Fisheries and Wildlife, 13 Natural Resources Bldg,

East Lansing, MI 48824, USA

Corresponding Author:

Alexander J. Jensen

Michigan State University, Department of Fisheries and Wildlife, 13 Natural Resources Bldg,

East Lansing, MI 48824, USA

(207) 659-2226 (Phone)

(517) 432-1699 (Fax)

jensen.alex1502@gmail.com

21 **Abstract**

22 A key uncertainty surrounding barrier removals in the Great Lakes is the response of
23 invasive sea lamprey populations to realized increases in available habitat for adfluvial species.
24 We addressed this uncertainty by applying a management strategy evaluation model, originally
25 developed to inform sea lamprey management in the Great Lakes, to forecast the effects of
26 barrier removal on Lake Michigan sea lamprey abundances. We used this model to characterize
27 the response to systematically increasing habitat availability and a specific proposed barrier
28 removal. Our results suggest the removals allow novel production from newly opened habitat
29 and, assuming a fixed budget for sea lamprey control, decrease the overall effectiveness of
30 control, leading to disproportionate increases in abundance. The case study demonstrated that
31 evaluating population effects only at the scale of watersheds directly affected by barrier removals
32 would substantially underestimate effects at the scale of Lake Michigan. Similar population
33 responses are possible when evaluating the effects on desired species. Our findings highlight the
34 importance of considering trade-offs for barrier removals and selecting the appropriate scale for
35 forecasting.

36 **Introduction**

37 Dams are ubiquitous features of watersheds throughout the world, and historically
38 provided many societal benefits, but they also serve as significant barriers to migratory fish.
39 Indeed, dams, hereafter referred to as stream barriers or simply barriers, have been implicated in
40 the declines of numerous populations of diadromous species (Limburg and Waldman 2009).
41 Thanks to growing public preference to increase lotic connectivity and benefit aquatic species,
42 barrier removal in the U.S. is accelerating and many large-scale structures have been demolished
43 in systems like the Penobscot, Carmel, and Elwha Rivers in Maine, California, and Washington,
44 respectively. Observed ecological benefits from previous barrier removals include increased
45 biological diversity, restoration of historical habitat, and enhanced passage (Bednarek 2001).
46 Ecological trade-offs emerge, however, when improved river access eliminates impediments to
47 the spread of unwanted species (McLaughlin et al. 2013). By restricting the range expansion of
48 invasive species, stream barriers in select systems may actually provide an important
49 conservation function by blocking fish migration (Sharov and Liebhold 1998; Vélez-Espino et al.
50 2011; Rahel 2013).

51 Sea lamprey have caused considerable ecological and economic damage within the
52 Laurentian Great Lakes since their invasion in the early 20th century (Smith and Tibbles 1980).
53 The parasitic juvenile stage of this species feeds on Great Lakes fish before maturing and
54 migrating to Great Lakes tributaries to spawn; the resulting larvae live as burrowing filter-
55 feeders in these streams for several years before metamorphosing and migrating back to the lakes
56 to begin their parasitic stage (Applegate 1950). Sea lamprey are currently controlled to generally
57 acceptable population levels in the Great Lakes using a combination of lamprey-specific
58 pesticide (*i.e.*, lampricide) applications and intentional fragmentation (Smith and Tibbles 1980).

59 A limited budget is allocated annually to both elements of control. Stream barriers play an
60 important role by preventing migratory adult sea lamprey from accessing high quality spawning
61 habitat, and consequently eliminating the need for costly treatments of large sections of rivers
62 (Hunn and Youngs 1980). The Great Lakes Fishery Commission (GLFC) Sea Lamprey Control
63 Program (SLCP) uses both pre-existing and actively constructed stream barriers to block sea
64 lamprey migration (Lavis et al. 2003).

65 In concert with the prospective benefits to resident fish species including various
66 salmonids and lake sturgeon (*Acipenser fulvescens*), Great Lakes barrier removals have the
67 potential to greatly reduce the effectiveness of sea lamprey control. In the Lake Michigan basin
68 alone, barriers like the Sixth Street Dam, Union Street Dam, and Calkins Bridge Dam currently
69 block hundreds of miles of viable spawning and larval habitat in the Grand River, Boardman
70 River, and Kalamazoo River, respectively. If these structures were removed without
71 construction of a replacement lamprey barrier or an increase in the lampricide control budget,
72 there would be two options available to control agents: 1) ignore production from the newly
73 available habitat, or 2) re-allocate lampricide application efforts to the newly available habitat as
74 needed, at the expense of reducing the frequency of applications in other river systems. The first
75 option is unlikely to be considered for large systems like the Grand River, while the second
76 option requires a shift in control effort from existing stream systems to the new habitat,
77 potentially decreasing treatment effectiveness across the basin as a whole. Although the
78 qualitative risks of barrier removal for sea lamprey management in the Great Lakes are accepted
79 by fishery management agencies, there is a need to better understand the actual magnitude of the
80 sea lamprey response to barrier removals.

81 In addition to the ecological concerns surrounding barrier removals, decision-makers are
82 faced with numerous competing objectives and pressures, including the maintenance of
83 infrastructure condition and public safety, generation of power, and enhancement of recreational
84 opportunities. The development of formal criteria, supported by the necessary scientific and
85 social information, is one solution for managing these trade-offs (Pejchar and Warner 2001).
86 With respect to the scientific information, researchers are specifically arguing for more careful,
87 comprehensive consideration of the potential ecological consequences and an increased role for
88 scientists in providing data on these consequences (Johnson and Graber 2002; Doyle et al. 2003).
89 In the case of barrier removals in the Great Lakes, research that equips managers with a more
90 explicit understanding of the effects of barrier removals on sea lamprey control can help
91 formalize the balancing of trade-offs inherent in decision-making.

92 Evaluating the expected effects of barrier removals requires consideration for the
93 appropriate spatial scale of modeling and relevant aspects of habitat quality upstream of barriers.
94 Most previous studies have focused on river-specific impacts of barrier removals (Stanley et al.
95 2007; Burroughs et al. 2010). The effects of barrier removals on sea lamprey populations,
96 however, are not restricted to river-specific production, as sea lamprey appear to exhibit a lack of
97 natal homing when migrating to tributaries to spawn (Bergstedt and Seelye 1995). Sea lamprey
98 production from a specific river can influence future spawner abundances in other rivers, so
99 predicting the effects of barrier removals on Great Lakes requires a consideration of population
100 dynamics on a larger scale than that of individual rivers. Furthermore, both the quality and
101 quantity of habitat upstream of barriers needs to be evaluated. Sea lamprey recruitment is known
102 to be limited by the availability of larval habitat, defined as substrate dominated by fine
103 sediments (Slade et al. 2003), and the attractiveness of river systems to migrating spawners is

104 partially driven by habitat quality and quantity (Morman et al. 1980; Mullett et al. 2003).
105 Previous modeling efforts looking at the effect of changing habitat availability on other fish
106 species have also emphasized the importance of habitat quality in predicting population
107 responses (Cheng et al. 2006; van der Lee and Koops 2016).

108 Management strategy evaluation (MSE) modeling, using known information about sea
109 lamprey life history and control in the Great Lakes, represents a feasible, realistic means to
110 capture the expected effects of barrier removals on the long-term effectiveness of sea lamprey
111 control. Management strategy evaluation models are powerful tools for research and
112 management because they tie together biological, observational, and management processes,
113 account for sources of uncertainty in each of these processes, and allow researchers to formally
114 compare the ability of competing management strategies to achieve specified management
115 objectives (Smith et al. 1999; Harwood and Stokes 2003). We have already developed an MSE
116 model for sea lamprey, specific to the Great Lakes, to assess the effect of alternative
117 management strategies (Jones et al. 2009). This model has been used to determine optimal
118 control budgets to achieve target economic injury levels (Irwin et al. 2012) and to explicitly
119 compare the effectiveness of alternative management strategies at a basin-wide scale (Dawson et
120 al. 2016).

121 We modified the MSE model to evaluate the effects of barrier removals on the Lake
122 Michigan sea lamprey population. Lake Michigan was selected as the focal spatial scale for this
123 work due to the observed lack of natal homing for sea lamprey within lakes and the detailed
124 understanding of sea lamprey population dynamics in this region (Dawson et al. 2016). We first
125 assessed the system's general response to increasing habitat availability through the incremental
126 addition of discrete habitat units with varying attributes of habitat quality. We also modeled a

127 specific Lake Michigan barrier removal scenario, using input data and management scenarios
128 defined by sea lamprey control agents, to inform decision-making for a contentious,
129 contemporary barrier removal scenario. Both approaches helped explain how a complex,
130 intensively managed biological system would respond to anthropogenic changes in habitat
131 availability.

132 **Methods**

133 *Model Description*

134 To evaluate the potential effect of barrier removals on sea lamprey production within an
135 MSE framework, we modified the MSE operating model developed by Jones et al. (2009) and
136 updated by Dawson et al. (2016). Briefly, this operating model includes interconnecting
137 biological, observational, and management components, operates at the spatial scale of an entire
138 Great Lake, and has an annual time step (Fig. 1). The biological model simulates the life history
139 of sea lamprey: adult sea lamprey from the lake habitat are allocated to streams for spawning;
140 these spawners produce stream-dwelling larvae according to a Ricker-type stock-recruitment
141 function; the larvae experience growth and mortality before metamorphosing into the parasitic
142 juvenile stages and migrating back to the lake. An observational model generates estimates of
143 stream-specific larval abundances intended to reflect measurement uncertainty with existing
144 sampling methods in the Great Lakes; these estimates are used to rank stream segments, called
145 treatment units, for treatment on the basis of cost per expected larva killed in the entire segment.
146 Treatment units are operationally defined as river sections treated with lampricides as a single
147 unit. The number of annually selected treatment units is limited by the total available control
148 budget. Treatment units selected for lampricide applications experience reductions in larval
149 abundance; the actual proportional reduction in abundance due to a lampricide treatment is
150 drawn from a beta distribution yielding average reductions of 93% and a CV of 0.10. Process
151 uncertainty is also included in the model in the form of a stochastic reproduction function
152 (Dawson and Jones 2009) and uncertainty in stream-specific larval growth rates. Further details
153 of the model's structure and parameterization are not repeated here; interested readers are
154 referred to earlier papers.

155 In addition to incorporating the capacity to flexibly add new habitat, as described below,
156 the model was altered to account for recent analyses of adult sea lamprey trapping data that re-
157 assessed the rules for allocating adult sea lamprey to spawning habitats. These modifications
158 included the following: 1) allocating 52% and 48% of all Lake Michigan spawners to northern
159 and southern tributaries, respectively, prior to allocating spawners to individual streams based on
160 drainage area and larval abundance, and 2) increasing the influence of drainage area, relative to
161 larval abundance, in determining spawner allocation to individual tributaries. Tributaries were
162 classified as northern or southern based on the location of their mouths relative to a dividing line
163 stretching across Lake Michigan from Frankfort, MI, to just south of Manistique, MI (Mullett et
164 al. 2003). These changes were made to match simulated spawner numbers with observed adult
165 distributions in sixteen Lake Michigan rivers that have received previous spawner assessments
166 (H. Dawson and M.L. Jones, Quantitative Fisheries Center, Michigan State University, East
167 Lansing, Michigan, unpublished analysis), and to reflect an updated analysis of sea lamprey
168 trapping data from throughout the Great Lakes that examined covariates affecting relative
169 spawning run size (Mullett et al. 2003; M.L. Jones, Quantitative Fisheries Center, Michigan State
170 University, East Lansing, Michigan, unpublished data).

171 *Population Responses to Systematic Barrier Removals*

172 We first characterized the general response of the Lake Michigan sea lamprey population
173 to barrier removals by systematically adding standardized habitat blocks. Each block was
174 assigned identical attributes, including areas of suitable larval sea lamprey habitat types as
175 defined by the GLFC (*i.e.*, Type I and Type II; Slade et al. 2003), drainage area, treatment cost,
176 and miscellaneous larval growth and mortality parameters; these are all attributes of existing
177 treatment units within the original operating model. Block attributes were calculated as averages

178 of all existing treatment units in Lake Michigan; each habitat block was assigned a total larval
179 habitat area of 386,275 m², drainage area of 842.8 km², and treatment cost of \$127,864. These
180 habitat additions were intended to simulate the effect of opening new river systems to sea
181 lamprey (*i.e.*, removing barriers at the river mouths).

182 The systematic addition of habitat was conducted in two ways: 1) combine new habitat
183 blocks into an ever larger single treatment unit or 2) add habitat blocks as multiple, discrete
184 treatment units. These two approaches were intended to contrast the effect of opening a single
185 large river with the effect of opening numerous small tributaries, with the same overall increase
186 in total habitat area. The single river is considered for treatment as a stand-alone system,
187 whereas each of the added small tributaries was ranked separately. When additional habitat
188 blocks were combined to form the single treatment unit, the total habitat area, drainage area, and
189 treatment cost were correspondingly increased in a 1:1 relationship; a treatment unit composed
190 of six habitat blocks would therefore have twice the drainage area, treatment cost, and habitat
191 area as one composed of three such blocks. We systematically assessed the effect of increased
192 habitat availability by adding three habitat blocks at a time. This was a convenient scale of
193 analysis because nine additional habitat units represent a 10% increase in total habitat
194 availability across Lake Michigan. In the end, we chose to evaluate increasing habitat
195 availability up to an additional 18 habitat units, representing a plausible range of changes in
196 overall habitat given existing barrier removal proposals in the Lake Michigan basin.

197 The influence of two categorical treatment unit attributes, namely recruitment potential
198 and geographically-determined spawner allocation, on the sea lamprey response to barrier
199 removals were formally evaluated by running increasing habitat addition simulations for each
200 possible combination of attributes. New habitat areas were either characterized as having high or

201 low recruitment potential, reflecting observed (Dawson et al. 2016) differences in Ricker stock-
 202 recruitment parameter estimates between streams classified by sea lamprey program control staff
 203 as regular versus irregular producers. Dawson et al. (2016) reported that recruitment potential
 204 (Ricker α estimates) was 3.4x greater in regular producers. Furthermore, habitat units were
 205 characterized as having elevated or reduced spawner allocation, based on whether they were
 206 assigned to northern or southern Lake Michigan, respectively. New habitat regions added to
 207 northern Lake Michigan were regarded as having elevated spawner allocations because 52% of
 208 all Lake Michigan spawners are assigned to this region, despite containing smaller rivers with
 209 smaller drainage areas and corresponding attractive flows for migrating sea lamprey compared to
 210 southern Lake Michigan.

211 For each removal scenario, we ran the model for 5 000 simulations, with a 100 year time
 212 horizon for each simulation; this was intended to capture the full range of stochasticity in model
 213 results and yield an equilibrium state for each simulation. For every simulation, the mean
 214 number of total lake-wide adult spawners across the last ten years was calculated to represent
 215 expected equilibrium conditions. The mean system response for each habitat addition scenario
 216 was summarized by calculating the percent change in mean abundance, across simulations, from
 217 status quo mean abundance using the equation below, in which the original value refers to mean
 218 status quo abundance unless otherwise stated:

219 (1)
$$\frac{(\text{New Value} - \text{Original Value})}{\text{Original Value}} \times 100$$

220 The simulated range of variation for each scenario represented variability among the
 221 simulation-specific 10-year averages. We also took advantage of the stochastic nature of the
 222 simulations to calculate the proportion of the 5 000 simulations, for each habitat addition
 223 scenario, exceeding a high threshold relative to average status quo spawner abundance; we

224 selected an abundance of 152 266 spawners based on the 90th percentile of simulated lamprey
225 abundances under status quo conditions. This simulated threshold abundance is similar to the
226 maximum estimated Lake Michigan adult abundance of 141 730 over a recent 10-year period
227 (2005-2014). Finally, to calibrate the model at the current Lake Michigan control budget of
228 \$2.42 million, larval survival was adjusted until the base model (*i.e.*, no habitat additions)
229 successfully projected the recently estimated average adult abundance of 72 200 (M. Siefkes,
230 Great Lakes Fishery Commission, Ann Arbor, Michigan, personal communication, 2016).

231 *Explaining Forecasted Population Trends*

232 To explain the forecasted trends in adult sea lamprey abundance with increasing habitat
233 availability, we ran additional simulations to characterize trends in the following model
234 components: stream-specific parasitic sea lamprey production, control budget allocation among
235 the newly added and original treatment units, and lampricide treatment frequency. Parasitic sea
236 lamprey production reflected the total number of metamorphosed sea lamprey leaving streams in
237 each year and simulation. Tracking stream-specific production facilitated comparison of the
238 relative contribution of the new and original treatment units to lake-wide adult abundances.
239 Additionally, looking at both control budget allocation and treatment frequency helped to explain
240 why the relative contributions of sea lamprey production from new and original treatment units
241 might change with increasing habitat availability.

242 We ran these additional simulations 1 000 times over the same 100 year timespan;
243 consistent with other simulations, only the last ten years of data in each simulation were used to
244 characterize trends. Simulations were run only for increasing habitat availability in which
245 regular producing streams were added to northern Lake Michigan, as these attributes produced
246 the strongest trends in sea lamprey abundance and were therefore more amenable for elucidating

247 population drivers. These simulations were run for the full range of increasing habitat
248 availability and for both the single large and multiple small river additions. We expect
249 qualitative patterns to be similar for other scenarios, such as simulating increasing habitat
250 availability in southern Lake Michigan streams.

251 *Case Study: Simulating A Barrier Removal on Michigan's Grand River*

252 We selected the potential removal of Michigan's Sixth Street Dam to demonstrate the
253 utility of an MSE approach in informing a potentially high impact barrier removal scenario. The
254 Sixth Street Dam is located in downtown Grand Rapids, MI, and has served as an important
255 incidental lamprey barrier on the Grand River, Michigan's longest river system. Approximately
256 96 river km lies between the Sixth Street Dam and the Webber Dam, the next upstream barrier
257 on the mainstem, and numerous large tributaries, including the Thornapple, Maple, and Rogue
258 Rivers drain into the Grand River between the two barriers, in addition to many smaller streams
259 (Fig. 2).

260 Recently, there has been pressure by citizen stakeholders to remove this barrier, with the
261 primary goals of recreating the historical rapids and establishing new recreational boating
262 opportunities (Adair and Sullivan 2015). Thanks in large part to the current relevance and
263 extent of currently protected upstream habitat, the Sixth Street Dam removal scenario was listed
264 a high priority for modeling by SLCP managers (P. Hrodey and M. Siefkes, Great Lakes Fishery
265 Commission, Ann Arbor, Michigan, personal communication, 2015). Furthermore, this system
266 can also be modeled with some degree of accuracy given the quantity of compiled data; SLCP
267 surveys for larval habitat quantities and native lamprey densities were conducted in 2014 and
268 2015, in addition to the recent development of treatment cost estimates for the area.

269 To simulate the removal of the Sixth Street Dam, we incorporated sixteen new treatment
270 units between the Sixth Street Dam and Webber Dam, each representing distinct Grand River
271 tributaries, into the model database. The mainstem of the Grand River was deemed likely to host
272 relatively low densities of larvae, thereby making treatment prohibitive from a cost-effective
273 standpoint (Fig. 2; J. Tews, U.S. Fish and Wildlife Service, Ludington, MI, personal
274 communication, 2015). Each included treatment unit was known to contain viable habitat for
275 spawning and larval sea lamprey, and had a uniquely estimable treatment cost. Additional
276 attributes of the new treatment units were then estimated using all available data on the Grand
277 River (supplementary data are available online).

278 Three primary management decisions were selected as the focus for modeling work: the
279 decision to modify the Webber Dam to block sea lamprey, the decision to treat or ignore the
280 newly available habitat upstream of the Sixth Street Dam, and the decision to maintain or
281 increase the current lake-wide control budget (Table 1). Because the Webber Dam currently has
282 the potential to pass sea lamprey but can be modified to block them, we simulated the influence
283 of barrier modification by allowing or denying sea lamprey access to the Looking Glass River;
284 this river is the only major tributary between the Webber Dam and the next mainstem barrier.
285 The decision to treat or ignore habitat upstream of the Sixth Street Dam was intended to compare
286 the effect of pulling treatment effort away from other Lake Michigan tributaries with the effect
287 of allowing uninhibited lamprey production above the Sixth Street Dam, respectively. Finally,
288 for the scenario in which the upstream system is treated and the Webber Dam blocks access to
289 the Looking Glass River, we both evaluated the effect of treating the system under the current
290 budget of \$2.42 million and estimated the necessary budget increase to prevent a lake-wide
291 increase in sea lamprey abundance above status quo levels.

292 We also formally assessed the influence of the assumed degree to which sea lamprey
293 utilize the newly available larval habitat upstream of the Sixth Street Dam. Among all inputs,
294 larval habitat quantity is especially important to evaluate given its observed role in influencing
295 recruitment success (Jones et al. 2003) and explicit incorporation into the operating model (Jones
296 et al. 2009). We therefore assessed the response of sea lamprey to two levels of assumed habitat
297 use within added tributaries for each of the control scenario combinations: 10% and 50% habitat
298 use. The 10% habitat use represents a reasonable approximation of expected lamprey use of total
299 river length based on professional judgment (A. Jubar, U.S. Fish and Wildlife Service,
300 Ludington, Michigan, personal communication, 2016) and preliminary analyses indicating that
301 the lengths of existing Grand River treatment units (obtained from the SLCP's database)
302 averaged just 10% of the total tributary lengths calculated from the GIS-based Sea Lamprey
303 Control Map (Great Lakes Fishery Commission 2016; A. Jensen, Michigan State University,
304 East Lansing, Michigan, unpublished analysis). Expected use of total river length is as low as
305 10% because linear referencing, in which even marginal lotic habitats unsuitable for larval sea
306 lamprey (*e.g.*, drainage ditches, ephemeral headwater creeks) are digitized to form stream GIS
307 datasets, can produce overestimates of total river lengths. We chose to assess the influence of
308 50% habitat use on the sea lamprey response in order to evaluate a presumed worst-case scenario
309 for extent of habitat use.

310 The model was run and summarized in the same manner as for the systematic habitat
311 additions (*i.e.*, 5 000 simulations, 100 year time horizon, ten year averages) for every scenario
312 and assumption, and the proportions of simulation results above the same status quo threshold
313 were again calculated.

314 **Results**

315 *Population Responses to Systematic Barrier Removals*

316 The simulated Lake Michigan sea lamprey population exhibited a nonlinear increase in
317 abundance in response to systematically increasing habitat availability that varied in magnitude
318 across the combinations of habitat addition attributes (Figs. 3, 4). The smallest percent increase
319 in mean abundance from status quo conditions with a 20% increase in habitat availability was
320 161%; the greatest increase exceeded 800%. The type of barrier removal (*i.e.*, whether there is
321 one large-scale barrier removal or multiple small-scale events) influenced the magnitude of the
322 sea lamprey population's response to barrier removal, with the addition of a single large stream
323 having the greater effect. The largest percent increase in abundance for the single stream
324 addition was 885%, compared to 452% for multiple stream additions. This difference in
325 abundance between the types of habitat addition held true across all combinations of recruitment
326 potential and spawner allocation. Corresponding with the different trends in mean abundance,
327 the proportion of simulations with forecasted abundances greater than the high threshold relative
328 to status quo abundance (152 266) also approached one more rapidly, relative to the amount of
329 added habitat, when additions were conducted as a single large river.

330 Whether the additional accessible habitat had high or low recruitment potential, as well as
331 whether it experienced high or low spawner allocation, also had implications for the simulated
332 effectiveness of sea lamprey control under barrier removal scenarios. Habitat additions with
333 high recruitment potential and high spawner allocation, which would correspond to habitat
334 assigned the status of regular producers and added to northern Lake Michigan, resulted in higher
335 abundances than habitat additions with low recruitment potential and low spawner allocation
336 (Figs. 3, 4). Between these two categorical factors, recruitment potential had the slightly greater

337 effect on resulting adult sea lamprey abundances. With a 20% increase in habitat availability and
338 the combination of spawner allocation and type of habitat addition held constant, high
339 recruitment habitat additions resulted in 38.2% to 115% greater mean adult sea lamprey
340 abundances relative to abundances arising from habitat additions with low recruitment potential.
341 With the same 20% increase in habitat availability, high spawner allocation habitat resulted in
342 mean abundances 23.3% to 92.2% greater than those achieved under habitat additions with low
343 spawner allocation.

344 *Explaining Forecasted Population Trends*

345 A combination of novel sea lamprey production from newly added habitat and increasing
346 production from the original treatment units, caused in part by a shifting allocation of treatment
347 effort away from original units to new ones, underlie the disproportionate response of adult sea
348 lamprey abundance to habitat increases. As expected, the average contribution of basin-wide sea
349 lamprey production from new treatment units increased in response to increasing absolute
350 amounts of new accessible habitat (Fig. 5a). Increasing habitat availability also caused a steep,
351 concurrent increase in production within the original treatment units (Fig. 5b); the nature of the
352 response was consistent across both types of habitat addition. This response may be explained in
353 part by the reduced overall annual treatment frequency among original treatment units with
354 increasing habitat additions (Fig. 5c). The average annual allocation of the control budget to
355 original treatment units declined from \$2.42 million to a median of \$2.07 and \$1.79 million for
356 the single and multiple treatment unit additions, respectively, when 18 new habitat blocks were
357 added to the Lake Michigan basin (Fig. 6).

358 *Case Study*

359 All management scenarios pertaining to the Sixth Street Dam removal forecasted large
360 increases in adult sea lamprey abundance in Lake Michigan, assuming the control budget
361 remains unchanged (Fig. 7). Among the simulations, the lowest mean percent increase in adult
362 abundance from status quo conditions was 52%. This occurred when the Webber Dam was
363 modified to block sea lamprey, new habitat units were treated, and sea lamprey used 10% of
364 available habitat. For the same scenario, just over 24% of simulations resulted in abundances
365 exceeding the status quo 90th percentile. The largest mean percent increase of 269% occurred
366 when an unmodified Webber Dam allowed sea lamprey to infest the Looking Glass River, none
367 of the new habitat units were treated, and sea lamprey used 50% of potentially available habitat.
368 Approximately 87% of simulations for this scenario resulted in spawner abundances exceeding
369 the status quo 90th percentile.

370 The decision to modify the Webber Dam, the decision to treat the upstream Grand River,
371 and the assumed degree of habitat use each had substantial effects on equilibrium sea lamprey
372 abundances, but the relative magnitude of effects differed. When the decision to treat and
373 assumed habitat use were otherwise held constant among scenarios, the percent difference in
374 mean lake-wide sea lamprey abundance between simulations including and excluding the
375 Looking Glass River ranged between 13.1% and 19.6%, with higher simulated abundances for
376 scenarios including the Looking Glass River. The decision whether or not to treat the upstream
377 Grand River system had a larger effect on sea lamprey numbers than the decision to modify
378 Webber Dam, with the decision to not treat these units resulting in a 40.4% to 52.1% increase in
379 average adult abundance. Assuming greater habitat utilization in the new treatment units had a
380 similarly large effect on equilibrium sea lamprey abundances (34.7% to 49.1% increase).

381 For the barrier removal scenario in which upstream habitat is treated and the Webber
382 Dam is modified to block sea lamprey, substantial increases in the annual Lake Michigan control
383 budget were needed to restore mean sea lamprey abundances to levels at or below status quo
384 under the two assumptions of habitat use. Simulations suggested an annual control budget of
385 \$2.62 million per year, representing a \$200 000 increase from the current budget, was needed to
386 maintain mean abundances at or just below status quo levels when assumed habitat use was 10%
387 (Fig. 8). A control budget of \$2.78 million was required when assumed habitat use was 50%,
388 representing an annual budget increase of \$360 000.

389 Discussion

390 The systematic habitat addition simulations showed that a heavily-controlled invasive
391 species, like sea lamprey, responds to the localized easing of key management-imposed
392 constraints in a disproportionate manner. The primary constraints on sea lamprey population
393 growth in the Great Lakes are habitat limitations created by barriers in large river systems and
394 lampricide treatment-induced mortality at the larval stage (Christie et al. 2003; Lavis et al. 2003).
395 When these two constraints were diminished by the addition of habitat and the subsequent
396 shifting of treatment efforts to these new habitat blocks, simulated sea lamprey production
397 increased in both the new and original river systems, leading to a large increase in forecasted
398 adult abundance. Similarly strong responses in population abundance to changing top-down
399 controls have been observed for mesopredators (*i.e.*, mesopredator release), where small
400 reductions in the abundance of apex predators trigger disproportionate increases in mesopredator
401 abundance (Ritchie and Johnson 2009). There is also evidence for sea lamprey of large
402 population responses to barrier failures in Lake Michigan: unrestricted colonization of 220 km of
403 the Manistique River above a degraded barrier in the late 1990s and early 2000s was associated
404 with approximately a 100% increase in the estimated Lake Michigan sea lamprey abundance
405 (Klar and Young 2004).

406 The forecasted disproportionate response can be explained in part by production of sea
407 lamprey from newly available habitat and in part by dilution of control intensity across the basin.
408 First, the simulated population increased due to an immediate contribution of sea lamprey
409 production from new habitats. Second, shifts in control effort allocation to include new
410 treatment units led to an overall simulated decrease in treatment frequency for the original
411 treatment units, which led to increased production, on average, from the these units.

412 We further hypothesize that the lack of density-dependent controls on this already
413 suppressed population compounded these shifts in treatment allocation and total sea lamprey
414 producing habitat by giving rise to a positive feedback effect. The sea lamprey population in
415 Lake Michigan has been reduced to abundances far below carrying capacity, defined at the lake-
416 level by limits on the abundance of available hosts; contemporary abundances are believed to be
417 at or below 10% of pre-control levels, and host abundances are much higher than they were at
418 the start of the control program. Consequently, the modeled population is not regulated by
419 density-dependent processes when management actions allow for increased recruitment except in
420 rare instances when large recruitment events trigger density-dependent compensation at the
421 individual stream level. In the near absence of density-dependent regulation, a positive feedback
422 cycle allows the population to rise to a carrying capacity defined by the estimated stream-level
423 stock-recruitment dynamics (Dawson and Jones 2009), subject to constraints imposed by a
424 density-independent lampricide control program. It is possible that the size to which the lamprey
425 population grows would be lower than forecasted in the more extreme scenarios modeled here,
426 constrained by host dynamics. The abundances would, nevertheless, be large enough to inflict
427 severe damage on host populations. Positive feedback effects have been predicted for other
428 fisheries systems under changing predation pressure (Kirby et al. 2009; Audzijonyte et al. 2013).

429 In total, the simulated new production from new habitats, increased production from old
430 habitats due to shifted control efforts, and the near absence of density-dependent compensation at
431 current sea lamprey abundance levels drove the large forecasted response in sea lamprey
432 abundance from a comparatively small increase in habitat. These results suggest that future
433 evaluations of barrier removals focusing on potential fish responses should consider broader
434 spatial scales, especially for systems in which species do not exhibit strict natal homing and

435 control effort is necessarily balanced among many streams. Without considering lake-wide
436 impacts of small-scale barrier removals, we would not have forecasted the disproportionate
437 population response.

438 Similar types of population responses to increased accessible spawning and rearing
439 habitat may occur for desirable fish species in the Great Lakes. Although many species relying
440 on nearshore or riverine habitat for spawning are known to exhibit homing behavior, increased
441 reproduction coupled with modest rates of straying from natal habitats could enhance future
442 reproductive success across broader spatial scales. Lake sturgeon and lake trout (*Salvelinus*
443 *namaycush*) were observed to exhibit overall straying rates of 0.105 and 0.60 in Lake Michigan,
444 while walleye (*Sander vitreus*) in Lake Erie exhibit moderate gene flow among populations
445 (Bronte et al. 2007; Strange and Stepien 2007; Homola et al. 2012). Although not assessed in the
446 Great Lakes, the straying rates of Chinook salmon (*Oncorhynchus tshawytscha*) ranged from 0.01
447 to 1.0 among spawning tributaries in Washington's Wenatchee River (Ford et al. 2015).

448 Our analysis also revealed that barrier removal decisions need to account for factors in
449 addition to habitat quantity to accurately assess the effects of barrier removal. The difference
450 between opening a single large river and multiple small river systems is due to the challenge of
451 incorporating increasingly expensive single-system treatments into the stream ranking system; if
452 there is insufficient budget remaining when a unit ranks for treatment, it will be passed over in
453 favor of lower ranked, less expensive systems. Supporting this, simulated trends in budget
454 expenditure and treatment frequency among original treatment units flatten with increasing
455 habitat availability only for the single large river addition (Figs. 5c, 6), while lamprey production
456 from this new habitat increases more steeply (Fig. 5a). Habitat attributes of recruitment potential
457 and spawner allocation, the latter associated with geographic location, also played important

458 roles in mediating the sea lamprey response to increasing barrier removals. Expected differences
459 among these habitat attribute scenarios may be mitigated by more flexible management strategies
460 (*e.g.*, based in part on professional judgment rather than a fixed algorithm) capable of accounting
461 for higher sea lamprey output from larger, more productive systems.

462 The high degree of variability within each of the barrier removal scenarios reflects very
463 real uncertainty in our understanding of sea lamprey dynamics and should be explicitly
464 recognized in decision making. One of the strengths of the MSE approach is the incorporation of
465 multiple sources of uncertainty (Bunnefeld et al. 2011); for our model, these sources included
466 stochasticity in biological processes, larval abundance assessments, and control efforts. The
467 resulting variability in model results implies that the forecasted mean responses in abundance are
468 by no means guaranteed outcomes. Instead, the results indicate a wide range of plausible
469 alternative outcomes. Reporting results as proportions of simulations with values above some
470 management-relevant threshold value demonstrates the likelihood of an undesirable outcome,
471 rather than simply focusing on a “best-guess”; decision-makers can use this information to assess
472 the risk of key decisions.

473 Simulation results for the removal of the Sixth Street Dam confirmed trends forecasted in
474 simulations of systematically increasing habitat availability. The case study also highlights the
475 importance of treating the upstream Grand River in the case of barrier removal. To ignore the
476 newly infested upstream habitat and continue a status quo treatment program resulted in
477 markedly higher sea lamprey abundances, despite the dilution of basin-wide treatment effort that
478 would have occurred if upstream habitat had been treated. The estimated increases in control
479 budget necessary to maintain sea lamprey at status quo abundances provide decision makers with
480 an estimate of the cost of a barrier removal. There are numerous other potential barrier removals

481 under consideration in the Great Lakes, including those in Lake Michigan's Boardman River and
482 Lake Superior's Black Sturgeon River, that could be evaluated with this MSE tool.

483 The case study simulations depended on several assumptions: that the Sixth Street Dam
484 will not be replaced by a seasonal barrier, that migrating sea lamprey will eventually utilize all
485 identified upstream tributary systems, and that the evaluated percent habitat use values (10%,
486 50%) bracket realistic values. Stakeholder groups have proposed the construction of a
487 seasonally-adjusted structure, in place of the Sixth Street Dam, to operate as a barrier only during
488 sea lamprey migrations (Adair and Sullivan 2015). We chose not to account for this possibility
489 in simulating the removal due to the uncertainty surrounding its actual installation and potential
490 success at blocking sea lamprey. If the goals of such a barrier are blocking sea lamprey and
491 allowing passage of other non-jumping, migratory species, the overlapping migration
492 phenologies of Great Lakes fish largely prevent the balancing of such objectives without
493 installation of an effective fishway (Vélez-Espino et al. 2011). Even partial barrier failures can
494 contribute to large increases in lake-wide sea lamprey abundances, as demonstrated by the
495 historical failure of a barrier on Michigan's Manistique River (Klar and Young 2004). The
496 assumption that sea lamprey can and will use all identified upstream tributaries for spawning has
497 been largely supported by previous barrier removal studies and our understanding of sea lamprey
498 migratory capacity. In coastal river systems smaller than the Grand River, sea lamprey have
499 been observed to quickly re-colonize previously blocked upstream habitat (Hogg et al. 2013;
500 Lasne et al. 2014). Sea lamprey also appeared to rapidly colonize upstream reaches of the
501 Manistique River in northern Michigan, a river section over 220 km in length, following the
502 partial failure of a blocking barrier, and access upstream tributary systems in Portugal's River
503 Mondego, a river system draining a watershed slightly less than half the size of Michigan's

504 Grand River (Almeida et al. 2002; Klar and Young 2004). Finally, we assumed sea lamprey
505 would likely use 10% of available river length in the upstream tributaries, and evaluated 50%
506 habitat use as a worst-case scenario. Although the 10% assumption can be considered a
507 reasonable estimate based on professional judgment and preliminary analyses, it remains a rough
508 approximation. Identifying reliable habitat area estimates in future modeling endeavors will
509 require more detailed GIS data integrating length and width information along streams, as well
510 as an improved empirical understanding of habitat use by sea lamprey within tributaries.

511 Other modeling-based approaches have been used to inform barrier removal decisions
512 and predict fish response to changing habitat availability, but none have matched both the extent
513 and resolution of our modeling efforts for sea lamprey populations. At the broadest extent,
514 barrier removal prioritization efforts synthesize multiple sources of information and strive to
515 optimize barrier removals across varying spatial extents like the Great Lakes or Pacific
516 Northwest, but often make simplifying assumptions in relating passability, stream length, and
517 habitat quality to future fish production (Zheng et al. 2009; Kemp and O’Hanley 2010; Moody et
518 al. 2017). At a smaller spatial extent, landscape models are increasingly used to predict indirect
519 aspects of fish response to barrier removal, like spawning success; these models often fail to
520 provide direct estimates of fish abundance (Steel et al. 2004; Spens et al. 2007). Finer resolution
521 modeling has also occurred to predict scenario-specific fish responses to individual barrier
522 removals using species-specific, population dynamics models for American shad (*Alosa*
523 *sapidissima*), walleye, and American eels (*Anguilla rostrata*); (McCleave 2001; Cheng et al.
524 2006; Harris and Hightower 2012).

525 To our knowledge, no previous studies have assessed fish responses to barrier removal
526 using a detailed, species-specific management strategy evaluation approach across a spatial

527 extent comparable to Lake Michigan, nor have they explicitly considered the implications of
528 barrier removals in a coupled management system with trade-offs. With sufficient demographic
529 information, this approach could be applied to other species of migratory fishes where
530 management is implemented at local scales but could potentially affect larger metapopulations.
531 Protection of ecosystems from invasive species and restoration of ecological connectivity in lotic
532 systems are two of the most important issues facing fishery managers in the Great Lakes and
533 elsewhere. Although focused on the former, the approach detailed in this report illustrates a tool
534 of potential utility for the challenge of reconciling these two issues.

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707 **Tables**

708 Table 1. Illustration of the simulated Grand River barrier removal scenarios (Fig. 7). The two
 709 habitat use alternatives (*i.e.*, 10%, 50%) were run for each of the four management scenarios
 710 below, while simulations evaluating the effect of an increasing control budget were run only for
 711 Scenario #1.

Scenario	Treat habitat above Sixth Street Dam?	Allow infestation of Looking Glass River?	Assumed habitat use
1	Yes	No	10%
			50%
2	No	No	10%
			50%
3	Yes	Yes	10%
			50%
4	No	Yes	10%
			50%

712

713 **Figure Captions**

714 **Figure 1.** Conceptual diagram for the sea lamprey MSE model. Solid and dashed lines indicate
715 component linkages within and among the individual biological, observational, and management
716 models, respectively.

717

718 **Figure 2.** Map of the Grand River mainstem (a) and the modeled Grand River system between
719 the Sixth Street Dam and North Lansing Dam (b). Only those tributaries in (b) identified as
720 “New Grand River Treatment Units” were explicitly considered in the simulations, and the
721 numbers correspond to numbered treatment unit names in Table S1. River flowline data were
722 obtained from the National Hydrography Dataset Plus (Bondelid et al. 2010) and state boundary
723 lines were obtained from ESRI and TomTom North America, Inc.

724

725 **Figure 3.** Adult sea lamprey abundance trends with increasing habitat availability, assuming
726 habitat is added within a single treatment unit. High and low spawner allocation and recruitment
727 potential refer to the assignment of streams as northern or southern streams and regular or
728 irregular producers, respectively. Boxes, whisker bars, and open circles represent the 25th and
729 75th, 10th and 90th, and 5th and 95th percentiles of simulated adult abundances, respectively.
730 Solid horizontal lines and black circles represent corresponding median and mean values,
731 respectively, and the gray squares indicate proportions of simulations with abundances greater
732 than the status quo 90th percentile.

733

734 **Figure 4.** Adult sea lamprey abundance trends with increasing habitat availability, assuming
735 habitat is added as independent treatment units. Boxes, whisker bars, and open circles represent

736 the 25th and 75th, 10th and 90th, and 5th and 95th percentiles of simulated adult abundances,
737 respectively. Solid horizontal lines and black circles represent corresponding median and mean
738 values, respectively, and the gray squares indicate proportions of simulations with abundances
739 greater than the status quo 90th percentile. The asterisk indicates mean lamprey abundance from
740 Scenario #1 of the Grand River case study, with an assumed 10% habitat use (see Figure 7).

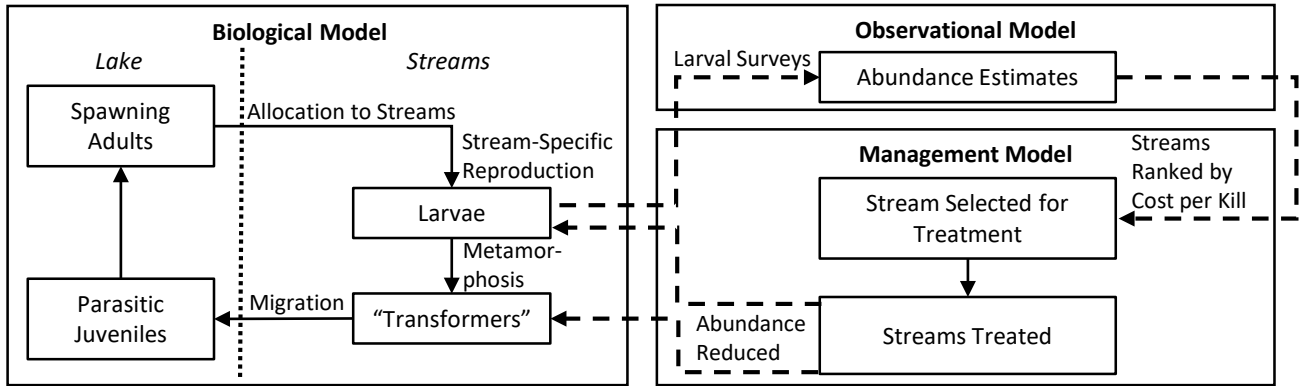
741
742 **Figure 5.** Changing model characteristics with increasing habitat. Lines and polygons represent
743 the median and 10th and 90th percentiles, respectively, across all simulations (a, b) or treatment
744 units (c). The dashed line and lighter polygon illustrate the effect of adding a single, large unit,
745 and the solid line and darker polygon illustrate the addition of habitat as multiple, discrete
746 treatment units, respectively.

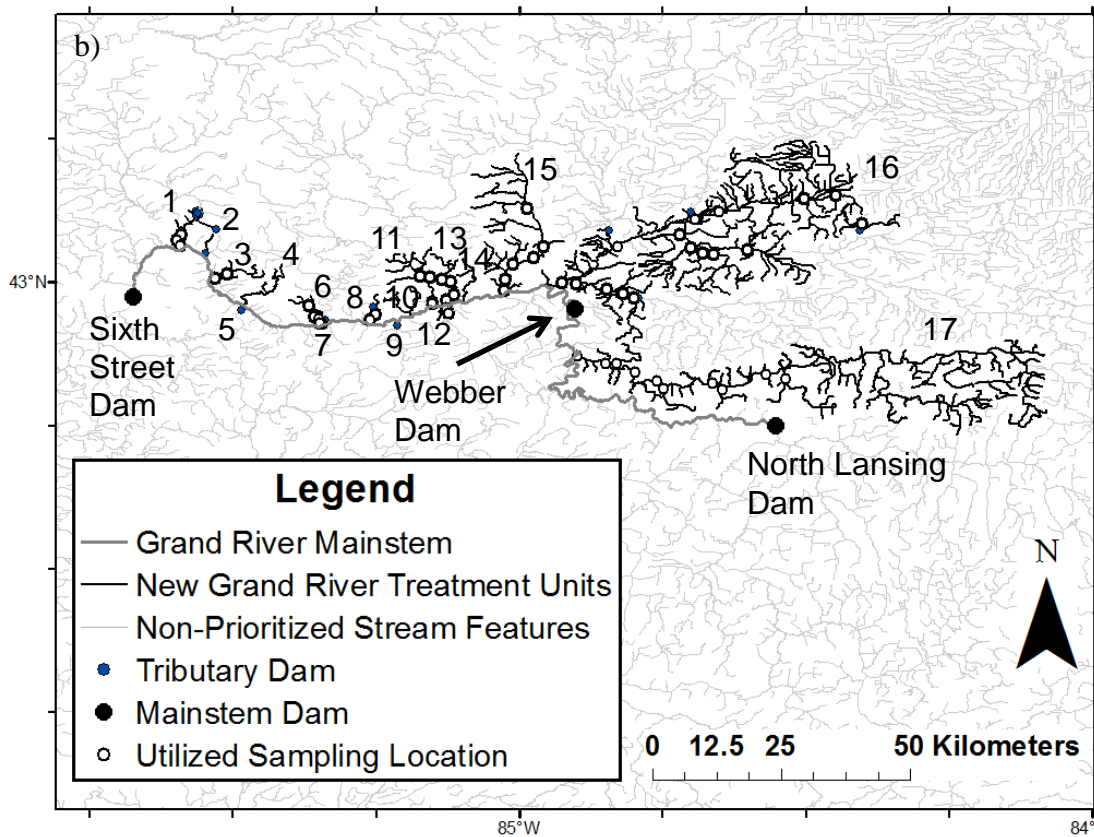
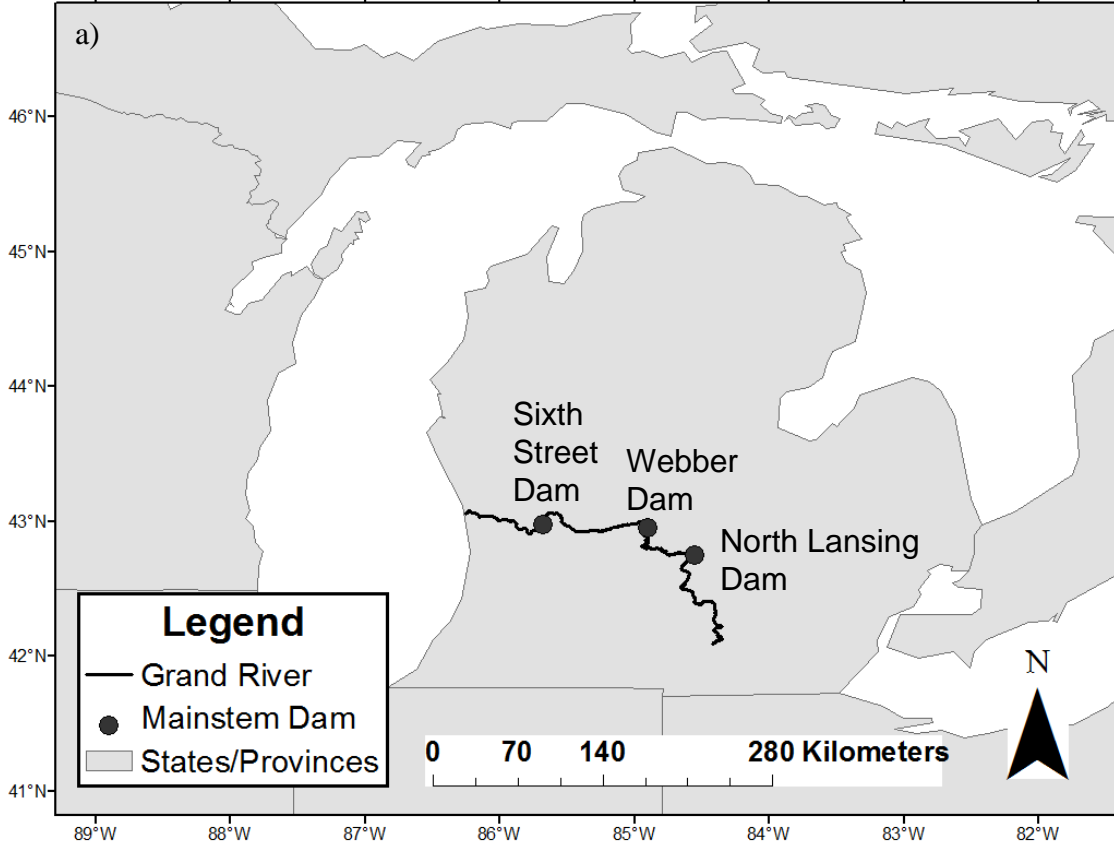
747
748 **Figure 6.** Average annual budget expenditure on the original treatment units with increasing
749 habitat availability. The dashed and solid lines illustrate the response when habitat is added as a
750 single, ever-larger system and multiple, discrete treatment units, respectively.

751
752 **Figure 7.** Expected sea lamprey abundances for each of the management scenarios. Scenarios
753 #1 and #2 exclude the Looking Glass River, while Scenarios #3 and #4 account for its influence.
754 New treatment units are treated by the SLCP in Scenarios #1 and #3, and ignored in Scenarios #2
755 and #4. Boxes, whisker bars, and open circles represent the 25th and 75th, 10th and 90th, and 5th
756 and 95th percentiles of simulated adult abundances, respectively. The solid horizontal lines and
757 black circles represent median and mean values, respectively. Numbers above the upper whisker
758 bars indicate the proportion of simulations greater than the status quo 90th percentile.

759

760 **Figure 8.** Expected sea lamprey abundances when the Sixth Street Dam is removed, the Webber
761 Dam is modified to block sea lamprey, lamprey are assumed to use 10% (a) or 50% (b) of
762 maximum potential river length, and the new treatment units are allocated control efforts with a
763 steadily increasing Lake Michigan control budget. Boxes, whisker bars, and open circles
764 represent the 25th and 75th, 10th and 90th, and 5th and 95th percentiles of simulated adult
765 abundances, respectively. The solid horizontal lines and black circles represent median and
766 mean values, respectively.

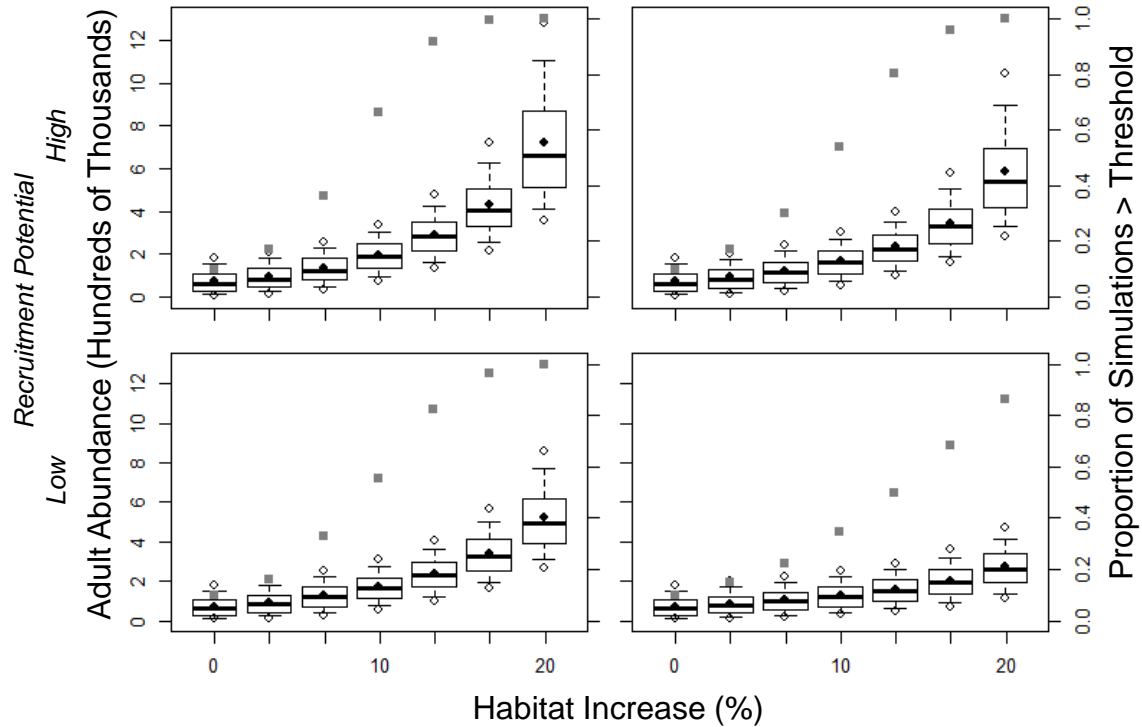




Spawner Allocation

High

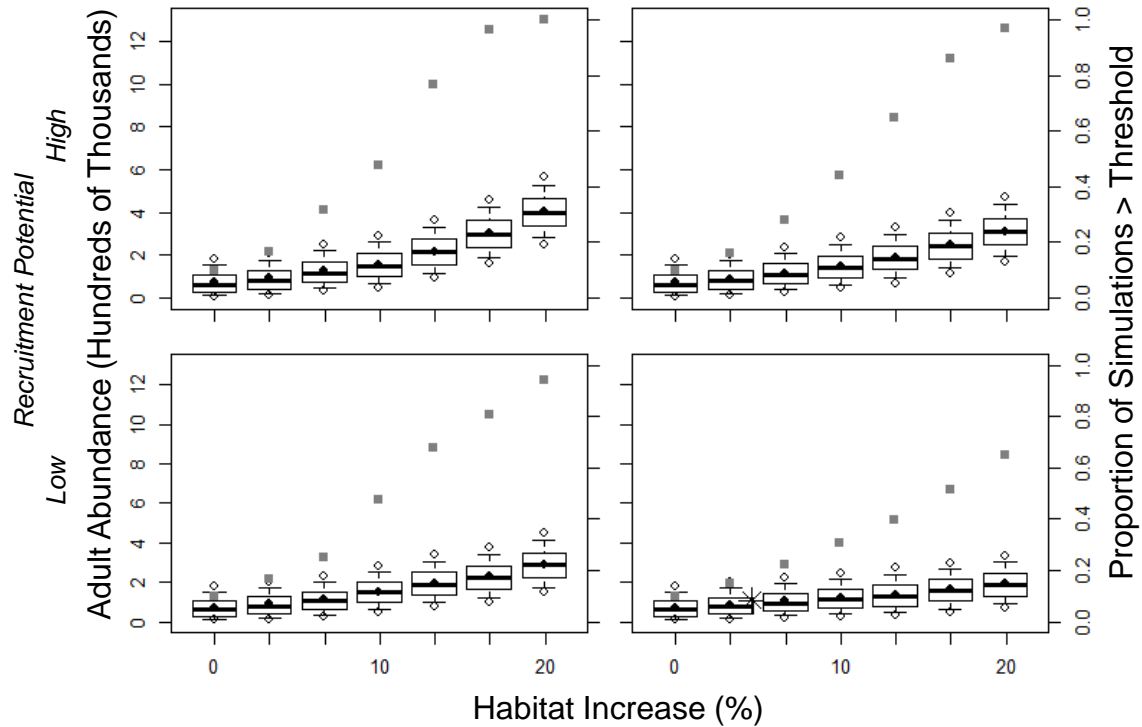
Low

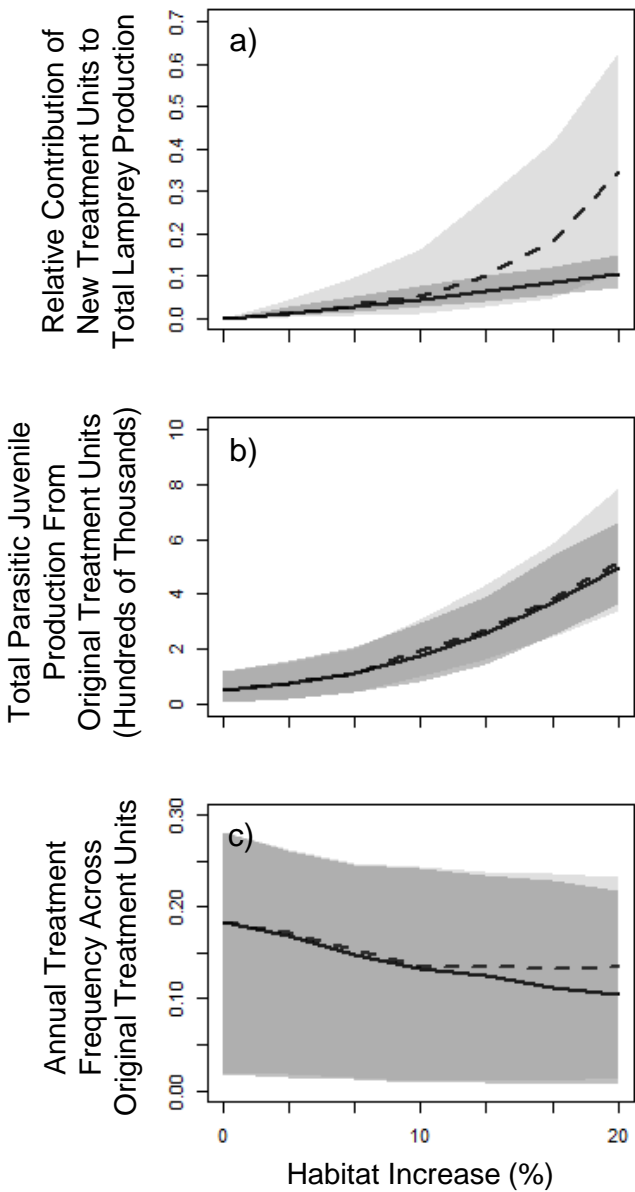


Spawner Allocation

High

Low





Budget Expenditure (\$ Million)
on Original Treatment Units

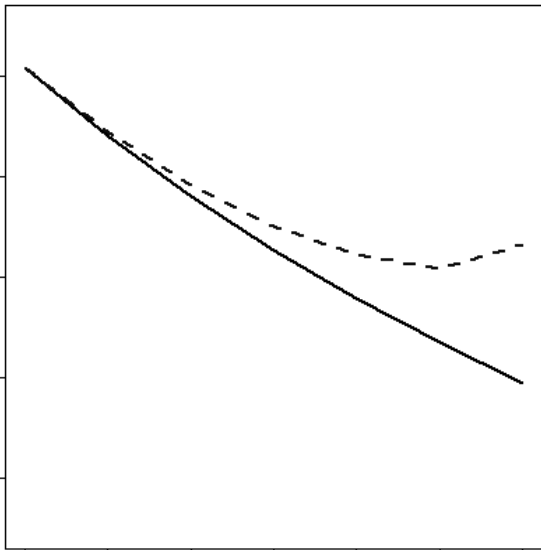
2.4
2.2
2.0
1.8
1.6

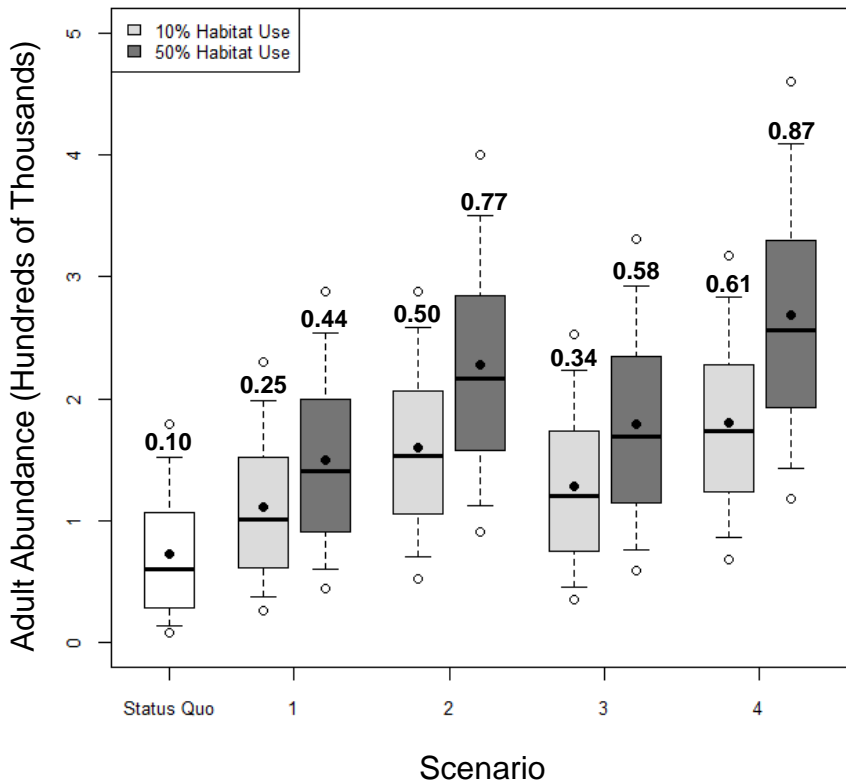
0

10

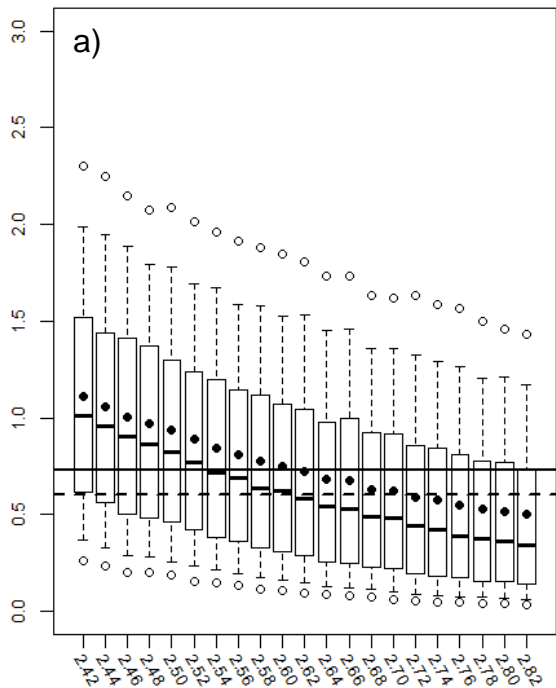
20

Habitat Increase (%)





Adult Abundance
(Hundreds of Thousands)



Control Budget (\$ Million)

